

Removal of Copper Toxicity by Zeolite in Java Tilapia *Oreochromis mossambicus* (Peters)

R. James, K. Sampath

Department of Zoology, V. O. Chidambaram College, Tuticorin 628 008, TN, India

Received: 21 October 2002/Accepted: 3 July 2003

Reduction of toxic elements in aquatic environments is one of the primary challenges in waste water treatment. The most widely used technique for removal of heavy metals involves the process of neutralization. Chemicals can effectively remove certain toxic substances from industrial wastes or polluted medium but it is presumed to be costly. However, there are some cheap chemicals which are also free from side effects. Zeolite is one such chemical and it acts as an ion-exchanging agent. Recently, it has been used in detergency, aquaculture ponds and nuclear treatment, but it also has large potential for other application in liquid waste treatment. Sodium aluminosilicate ($\text{Na}_{12}[(\text{AlO}_2)_{12}(\text{SiO}_2)_{12}]\cdot 27\text{H}_2\text{O}$) is one of the natural zeolites (faujasite). Previous authors have studied the role of zeolite in reducing toxic metabolites, settlement of suspended solids and absorption of gases like CO_2 , SO_2 and H_2S in aquaculture ponds (Chaberbain 1988; Chien 1992; Briggs and Smith 1996) and improvement of sediment quality, pH (Clifford 1992) and mineral nutrition in fish and shrimp (Battes *et al.* 1981). However, there is paucity of information on how zeolite reduces the toxic elements in polluted medium. There is also not much information available on the duration of the treatment and optimum dosage of zeolite needed to reduce the metal toxicity. The present work was designed to study the effect of zeolite on the removal of copper at two different levels (sublethal and median lethal) from water and tissues of Java tilapia, *Oreochromis mossambicus*.

MATERIALS AND METHODS

Oreochromis mossambicus (Peters) were collected from a pond and acclimated for 30 d in laboratory conditions. During acclimatization, water was changed daily and fish were fed *ad libitum* with pellet feed containing 35% protein. Three series of experiments were performed.

In the first series of experiments, 96 h LC_{50} value was determined following the static renewable bioassay method (Sprague 1973). Fish were starved for 24 h prior to the experiment and throughout the bioassay study. Stock solution of copper was prepared by dissolving 3.93 g of analar grade copper sulphate ($\text{CuSO}_4\cdot 7\text{H}_2\text{O}$) in 1 l of distilled water and then diluted with

freshwater to obtain the desired concentrations of the medium. Well acclimated and active 10 fish (12.4 ± 1.1 g) were exposed to different concentrations of copper (0, 3.0, 3.5, 4.0, 4.5, 5.0, 5.5, 6.0 and 6.5 mg l⁻¹) and mortality was observed for 96 h. A control was run in freshwater. The LC₅₀ was calculated adopting probit analysis. The 96 h LC₅₀ value of copper for *Oreochromis mossambicus* was 4.27 mg l⁻¹ and its 95% confidence limits were 3.72 (lower) and 4.90 (upper) mg l⁻¹.

In the second and third series of experiments, animals were exposed for 180 d to estimate metal accumulation in different tissues of fish and also metal content was determined in the medium and sediment. Acclimated animals (12.4 ± 1.1 g) recruited from stock were divided into six groups of 60 individuals each. They were exposed to the sublethal level of copper (2.14 mg l⁻¹) alone and with different levels of zeolite (series 2). Similarly another set of six groups of 60 individuals each was prepared and exposed to median lethal level of copper (4.27 mg l⁻¹) alone and alongwith different levels of zeolite (series 3). The animals were starved for 24 h before commencement of the experiment.

Table 1. Experimental groups and their notations.

S.No.	Groups	Total amount of zeolite added (kg)	Notation
<i>Sublethal treatments (SL)</i>			
1	Control (Freshwater)	—	C
2	Copper (2.14 mg l ⁻¹) alone	—	SLCu
3	Copper (2.14 mg l ⁻¹) + 0.5gZ*l ⁻¹	0.650	SLCuZ1
4	Copper (2.14 mg l ⁻¹) + 2.0gZ*l ⁻¹	2.600	SLCuZ2
5	Copper (2.14 mg l ⁻¹) + 4.0gZ*l ⁻¹	5.200	SLCuZ3
6	Copper (2.14 mg l ⁻¹) + 8.0gZ*l ⁻¹	10.400	SLCuZ4
<i>Median lethal treatments (ML)</i>			
1	Control (Freshwater)	—	C
2	Copper (4.27 mg l ⁻¹) alone	—	MLCu
3	Copper (4.27 mg l ⁻¹) + 0.5gZ*l ⁻¹	0.650	MLCuZ1
4	Copper (4.27 mg l ⁻¹) + 2.0gZ*l ⁻¹	2.600	MLCuZ2
5	Copper (4.27 mg l ⁻¹) + 4.0gZ*l ⁻¹	5.200	MLCuZ3
6	Copper (4.27 mg l ⁻¹) + 8.0gZ*l ⁻¹	10.400	MLCuZ4
* Sodium aluminosilicate			

For convenient presentation, control and experimental groups hereafter would be referred by their notation. The chosen levels of zeolite and a constant level of copper [2.14 mg l⁻¹ (50% of LC₅₀) in the case of sublethal exposure and 4.27 mg l⁻¹ in the case of median lethal exposure] were added to the medium separately on 0 d. The experiment was conducted in epoxy coated cement tank containing 1300 l of test medium. Pond sediment was added to

all the experimental tanks upto 5 cm height to simulate field condition and then 1300 l of freshwater was pumped in it. The medium was mixed well after the addition of copper and zeolite and then test animals were introduced. Zeolite treated fish did not show any deleterious effects but it improved the growth and haematological parameters during the experimental period of 45 days in fishes (James and Sampath 1999 a and b, James 2000), hence, individual effect of zeolite was not studied in the present study. 35% protein feed was offered as diet to test animals in a feeding tray once in a day at 0800 h and uneaten food was removed 2 h after feeding. The medium was not changed during the experiment and the medium was aerated for 14 h a day to avoid depletion of oxygen. The hydrobiological parameters like dissolved oxygen, temperature, pH, salinity and hardness of media were estimated during non-aeration and these parameters averaged to 4.16 ± 0.2 or 4.07 ± 0.4 ml l⁻¹, 29.2 ± 1 or $29.0 \pm 1^\circ\text{C}$, 7.80 ± 0.08 or 7.82 ± 0.08 , 0.15 ± 0.02 or 0.14 ± 0.02 g⁻¹ and 58.0 ± 4.0 or 60.0 ± 3.0 mg CaCO₃l⁻¹ in sublethal or median lethal exposures respectively. First week of every month, eight fish were removed from each group for estimation of metal accumulation in tissues gill, liver and muscle following the method of FAO (1975). There was no mortality in all experimental groups except ^{ML}Cu which recorded 16.7% mortality. The zeolite deposited over the sediment was carefully removed prior to sampling of sediment and dried at 60°C in hot air oven for 2 days. Copper concentration in water and sediment was estimated using the method of APHA (1993) and Smith and Windom (1972). All samples were analyzed in an Atomic Absorption Spectrophotometer (GBC 906 AA model). The instrument was calibrated using standards prepared from copper sulphate. Students 't' test was used to determine the significance of mean values between control and experimental groups at different days.

RESULTS AND DISCUSSION

Trace level of Cu was found in control water (0.005 mg l⁻¹) and sediment (0.024 mg l⁻¹) in the present study. The initial level of Cu in all exposures was 2.14 or 4.27 mg l⁻¹ in sublethal or median lethal exposures and it gradually declined with extension of exposure period from 0 to 180 days (Fig.1 and 2). When zeolite was added to sublethal level of Cu media, the metal content was completely removed within 120 days in ^{SL}CuZ2 - ^{SL}CuZ4 groups and 150 days in ^{SL}CuZ1 group. However, in median lethal exposures, Cu was completely removed on 180 days in ^{ML}CuZ2 - ^{ML}CuZ4 groups (except ^{ML}CuZ1). This indicates that zeolite took 60 more days for the complete removal of Cu in median lethal exposures as compared to sublethal exposures. The Cu level in ^{SL}Cu and ^{ML}Cu groups was 0.850 and 1.39 mg l⁻¹ even at the time of termination of the experiment. The Cu level in sediment and medium was inversely proportional (Fig.1 and 2).

Trace level of copper was also found in tissues of control fish and the amount of Cu had linearly increased with extension of exposure period in

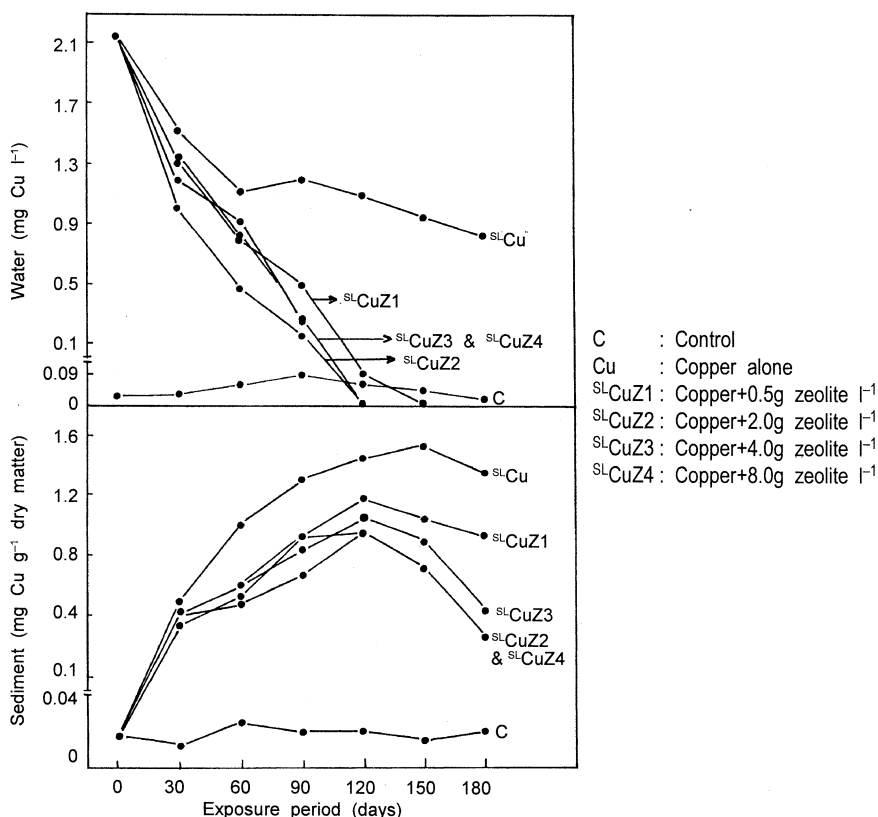


Figure 1. Effect of sublethal concentration of copper (2.14 mg l⁻¹) and addition of different doses of zeolite on copper distribution in water and sediment.

fish exposed to sublethal and median lethal levels of Cu alone. However, in zeolite treatments, there was initial increase in Cu content and the accumulation attained maximum on days 30 and 60 (except SL-CuZ1 and ML-CuZ1) and afterwards it drastically declined and completely eliminated from tested tissues on days 120 and 180 in sublethal and median lethal exposures respectively (Table 1 and 2). Fish exposed to median lethal exposures took 60 days more for the complete removal of Cu from tissues than those exposed to sublethal level.

Briggs and Smith (1996) found that zeolite have the capacity to remove ammonia and other metabolites from freshwater by ion-exchange and adsorption. James *et al.* (1998) reported that addition of another chelating agent EDTA to Cu contaminated medium cause the formation of stable ion (Cu²⁺) exchanged EDTA complex and elimination of more amount of copper in faeces, which significantly reduced the metal burden in tissues and improved the haematological parameters in *Oreochromis mossambicus*. Muramota (1980) found that metal chelating compounds NTA (Nitrilotriacetic acid) and EDTA reduced the metal toxicity in fish by preventing the accumulation of metal in tissues. He also

Table 2. Effect of sublethal concentration of copper (2.14 mg l⁻¹) and addition of zeolite on the copper accumulation (µg.g⁻¹ wet tissue) in tissues of *Oreochromis mossambicus*.

Expos- ures	Exposure period (days)						
	0	30	60	90	120	150	180
Gill							
C	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0
SL Cu	0.01 ± 0	1.51 ± 0.02	2.63 ± 0.14	4.65 ± 0.31	5.90 ± 0.12	8.17 ± 0.41	10.07 ± 0.26
SL CuZ1	0.01 ± 0	0.88 ± 0.01	0.33 ± 0.02	0.09 ± 0.16	0.07 ± 0	CuNd	CuNd
SL CuZ2	0.01 ± 0	0.27 ± 0.01	0.20 ± 0.01	0.02 ± 0	CuNd	CuNd	CuNd
SL CuZ3	0.01 ± 0	0.31 ± 0.02	0.11 ± 0.01	0.02 ± 0	CuNd	CuNd	CuNd
SL CuZ4	0.01 ± 0	0.40 ± 0.03	0.13 ± 0.02	0.02 ± 0	CuNd	CuNd	CuNd
Liver							
C	0.05 ± 0.01	0.05 ± 0	0.04 ± 0	0.06 ± 0	0.04 ± 0	0.05 ± 0	0.05 ± 0
SL Cu	0.05 ± 0.01	1.98 ± 0.02	3.25 ± 0.04	8.63 ± 0.26	12.74 ± 1.31	18.11 ± 1.26	21.35 ± 2.08
SL CuZ1	0.05 ± 0.01	0.92 ± 0.11	0.98 ± 0.06	0.13 ± 0.02	0.07 ± 0	CuNd	CuNd
SL CuZ2	0.05 ± 0.01	0.98 ± 0.06	0.62 ± 0.06	0.12 ± 0.03	0.01 ± 0	CuNd	CuNd
SL CuZ3	0.05 ± 0.01	0.99 ± 0.11	0.73 ± 0.03	0.15 ± 0.04	0.03 ± 0	CuNd	CuNd
SL CuZ4	0.05 ± 0.01	1.06 ± 0.13	0.85 ± 0.05	0.17 ± 0.01	0.02 ± 0	CuNd	CuNd
Muscle							
C	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0
SL Cu	0.01 ± 0	0.72 ± 0.06	1.85 ± 0.02	2.97 ± 0.04	4.06 ± 0.06	5.17 ± 0.14	4.23 ± 0.23
SL CuZ1	0.01 ± 0	0.15 ± 0.01	0.11 ± 0	0.01 ± 0	CuNd	CuNd	CuNd
SL CuZ2	0.01 ± 0	0.06 ± 0	0.03 ± 0	CuNd	CuNd	CuNd	CuNd
SL CuZ3	0.01 ± 0	0.10 ± 0	0.10 ± 0	CuNd	CuNd	CuNd	CuNd
SL CuZ4	0.01 ± 0	0.09 ± 0	0.25 ± 0.01	0.03 ± 0	CuNd	CuNd	CuNd

Each value is the mean ($\bar{X} \pm SD$) of three fish; CuNd: Copper not detected.

Table 3. Effect of median lethal concentration of copper (4.27 mg l⁻¹) and addition of zeolite on metal accumulation (µg/g wet tissue) in tissues of *Oreochromis mossambicus*.

Expos- ures	Exposure period (days)						
	0	30	60	90	120	150	180
Gill							
C	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0
ML Cu	0.01 ± 0	4.87 ± 0.31	9.62 ± 0.71	19.26 ± 1.16	23.33 ± 1.65	28.11 ± 2.16	29.77 ± 2.74
ML CuZ1	0.01 ± 0	2.45 ± 0.11	4.84 ± 0.26	6.80 ± 0.61	8.34 ± 0.48	7.74 ± 0.44	5.83 ± 0.36
ML CuZ2	0.01 ± 0	1.36 ± 0.03	2.43 ± 0.08	1.10 ± 0.06	0.70 ± 0.06	0.09 ± 0	CuNd
ML CuZ3	0.01 ± 0	1.40 ± 0.01	2.87 ± 0.12	1.42 ± 0.10	0.92 ± 0.03	0.06 ± 0	CuNd
ML CuZ4	0.01 ± 0	1.68 ± 0.13	2.27 ± 0.16	1.51 ± 0.09	1.13 ± 0.07	0.06 ± 0	CuNd
Liver							
C	0.05 ± 0	0.05 ± 0	0.05 ± 0	0.06 ± 0	0.06 ± 0	0.06 ± 0	0.05 ± 0
ML Cu	0.05 ± 0	7.74 ± 0.15	18.97 ± 1.36	29.44 ± 2.26	36.17 ± 3.51	43.43 ± 3.08	45.86 ± 4.11
ML CuZ1	0.05 ± 0	2.14 ± 0.11	6.90 ± 0.14	7.09 ± 0.19	9.24 ± 1.10	10.75 ± 0.93	8.26 ± 0.76
ML CuZ2	0.05 ± 0	2.81 ± 0.06	5.26 ± 0.28	4.16 ± 0.21	2.21 ± 0.01	0.60 ± 0	CuNd
ML CuZ3	0.05 ± 0	3.76 ± 0.17	6.91 ± 0.16	5.83 ± 0.28	2.96 ± 0.14	0.84 ± 0.01	CuNd
ML CuZ4	0.05 ± 0	1.89 ± 0.04	7.44 ± 0.31	7.84 ± 0.31	3.21 ± 0.30	0.87 ± 0.01	CuNd
Muscle							
C	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0	0.01 ± 0
ML Cu	0.01 ± 0	1.31 ± 0.02	3.03 ± 0.24	4.10 ± 0.33	6.37 ± 0.48	9.39 ± 0.75	11.18 ± 0.97
ML CuZ1	0.01 ± 0	0.63 ± 0.01	1.28 ± 0.02	1.93 ± 0.01	2.60 ± 0.06	2.05 ± 0.03	0.063 ± 0
ML CuZ2	0.01 ± 0	0.51 ± 0.03	0.73 ± 0	0.60 ± 0	0.27 ± 0	0.11 ± 0	CuNd
ML CuZ3	0.01 ± 0	0.62 ± 0.07	0.97 ± 0.01	0.82 ± 0	0.46 ± 0	0.04 ± 0	CuNd
ML CuZ4	0.01 ± 0	0.74 ± 0.05	1.23 ± 0.03	0.93 ± 0.01	0.53 ± 0	0.08 ± 0	CuNd

Each value is the mean ($\bar{X} \pm SD$) of three fish; CuNd: Copper not detected.

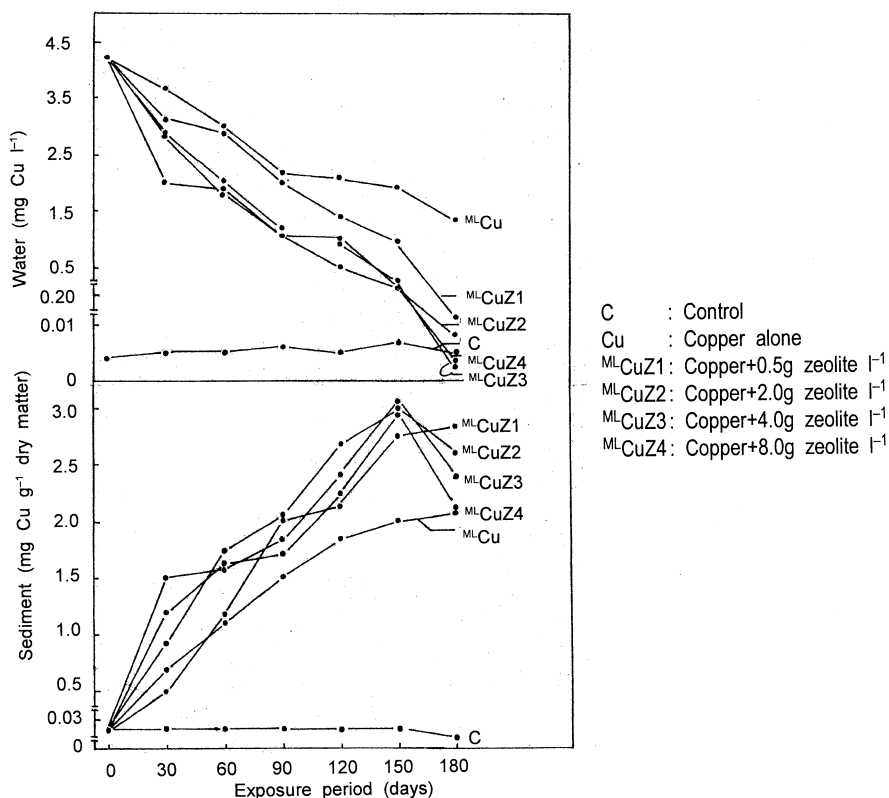


Figure 2. Effect of median lethal concentration of copper (4.27 mg l^{-1}) and addition of different doses of zeolite on copper distribution in water and sediment.

suggested that cadmium complexed to EDTA is indeed taken up, but the complex is quickly excreted through urine (Babiker and Rankin 1975).

The present study revealed that metal uptake by test animals was lower in zeolite treated groups than in the Cu exposed group. It was found that zeolite caused the elimination of Cu in polluted water (Fig.1 and 2) which in turn reduced the metal burden in fish (see Table 1 and 2). Among the zeolite treated groups, $^{SL}\text{CuZ2}$ in sublethal exposures and $^{ML}\text{CuZ2}$ in median lethal exposures (2 g Z l^{-1}) have elicited the better performances than other groups ($0.5, 4$ and 8 g Z l^{-1}) and hence an addition of $2 \text{ g zeolite l}^{-1}$ is considered as optimum dose. Based on the present study, it is recommended that, application of $2 \text{ g zeolite l}^{-1}$ to metal polluted environment could remove the heavy metal, copper at any level of concentration and reduce its toxicity on fish and other commercially important organisms. James and Sampath (1999b) reported that addition of $1 \text{ g zeolite l}^{-1}$ to cadmium polluted environment, resulted in least accumulation of metal (Cd) and maximum improvement in food intake and growth in the cadmium exposed

Heteropneustes fossilis. Gworek and Borowiak (1990) reported that application of synthetic zeolite cause the immobilization of heavy metals and recommended for clean up method. In aquaculture practices, application of zeolite is suggested in ponds before stocking fry or during the pond preparation (Briggs and Smith 1996). When the concentration of ammonium ion exceeds the permissible limit in aquaculture ponds, it becomes toxic to fish life (Sampath *et al.* 1991; James *et al.* 1993); hence it is advisable to reduce the concentration below the permissible limit by application of zeolite. However, doses of chelating agent like EDTA could cause deleterious effects on survival, development and growth in crustacean larvae (Simon 1981) and on survival and haematology of fish (James *et al.* 1998). Comparatively, zeolite is the cheapest, causes no side effects and more suitable than EDTA and NTA and hence it may be considered as one of the best chemical agents to remove toxic elements from polluted environments.

Acknowledgment. Financial assistance from TNSCST, Tamilnadu, India (TNSCST/STP/VR/ESO 3013/95-96) to Dr. R. James is gratefully acknowledged.

REFERENCES

- APHA (1993) Standard methods for examination of water and waste water (American Public Health Association, Washington DC)
- Babiker MM, Rankin JC (1975) Rationable for the use 51 Cr EDTA for estimation of glomerular filtration rate in fish. *Comp Biochem Physiol* 50A: 177-179
- Battes K, Marton A, Apertoraiei M, Rusu MA, Minciu D, Abraham A, Bucur N, Cachita D (1981) The utilization of mineral zeolites and of a biostimulator (procaine) as additives in concentrated fodder in the controlled culture of the carp. *Bull Cerc Piscicult Series Noua* 3: 45-58
- Briggs MRP, Smith SJF (1996) The effects of zeolite and other aluminosilicate clays on water quality at various salinities. *Aquat Res* 27: 301-311
- Chamberlain G (1988) Pond bottom degradation, *Coast Aquacult* 5: 2-5
- Chien YH (1992) Water quality requirements and management for marine shrimp culture. In: *Proceedings of the Special Session on Shrimp Farming*, Wyban J (Ed.), World Aquaculture Society, Baton Rouge, LA, p.144-156
- Clifford HC (1992) Marine shrimp pond management: a review. In: *Proceedings of the Special Session on Shrimp Farming*, Wyban J (Ed.), World Aquaculture Society, Baton Rouge, LA, Pp.110-137
- FAO (1975) Manual of methods in aquatic environment research. Part I, Publ. Div. FAO, Rome Pp.223
- Gworek B, Borowiak M (1990) Zg 1 Pat No. 283821 Poland (Patent notification)

- Huheey JE (1983) Inorganic chemistry. Principles of structure and reactivity, Harper International SI edition, London, Pp.73
- James R (2000) Effect of zeolite on reduction of cadmium level in water and improvement of haematological parameters in *Oreochromis mossambicus* (Peters). Indian J Fish 47: 29-35
- James R, Sampath K (1999a) Effect of zeolite on the reduction of cadmium toxicity in water and a freshwater fish, *Oreochromis mossambicus*. Bull Environ Contam Toxicol 62: 222-229
- James R, Sampath K (1999b) Effect of zeolite on reduction of cadmium toxicity: An experimental study on element uptake and growth in *Heteropneustes fossilis* (Bloch). J Aqua Trop 14: 65-74
- James R, Sampath K, Narayanan M (1993) Effect of sublethal concentration of ammonia on food intake and growth in *Mystus vittatus*. J Environ Biol 14: 243-248
- James R, Sampath K, Selvamani P (1998) Effect of EDTA on reduction of copper toxicity in *Oreochromis mossambicus*. Bull Environ Contam Toxicol 60: 487-493
- Muramota S (1980) Decreases in cadmium concentration in Cd contaminated fish by short term exposure to EDTA. Bull Environ Contam Toxicol 25: 828-831
- Sampath K, Sivakumar V, Sakthivel M, James R (1991) Lethal and sublethal effects of ammonia on survival and food utilization in *Oreochromis mossambicus* (Pisces, Cichlidae) J Aqua Trop 6: 223-230
- Sanderson RJ (1960) In: Chemical and periodicity, Van Nostrand Reinhold Company, New York, Pp.42
- Simon CM (1981) Design and operations of a large scale commercial penaeid shrimp hatchery. J World Maricult Soc 12:322-334
- Smith GR, Windom HL (1972) Analytical handbook for the determination of As, Cd, Co, Cu, Fe, Pb, Mn, Hg, Ni, Ag and Zn in the marine and estuarine environments. In: The technical report series of the marine science programs, University system of Georgia, No:72-76, Pp.62
- Sprague JB (1973) The ABC's of pollutant bioassay using fish. In Biological Methods for the Assessment of Water Quality. ASTM. STP 528 American Soc Testing Materials, p.6-30